

Risk of intact forest landscape loss goes beyond global agricultural supply chains

Siyi Kan ^{1,6}, Bin Chen ^{2,10*}, U. Martin Persson ³, Guoqian Chen ^{1,4*}, Yutao Wang ²,
Jiashuo Li ⁵, Jing Meng ⁶, Heran Zheng ⁶, Lan Yang ², Rui Li ⁷, Mingxi Du ⁸, Thomas
Kastner ⁹

¹ Laboratory of Systems Ecology and Sustainability Science, College of Engineering, Peking University, Beijing 100871, China

² Fudan Tyndall Center, Department of Environmental Science & Engineering, Fudan University, Shanghai 200438, China

³ Department of Space, Earth & Environment, Physical Resource Theory, Chalmers University of Technology, Gothenburg 41296, Sweden

⁴ Macao Environmental Research Institute, Macau University of Science and Technology, Macao 999078, China

⁵ Institute of Blue and Green Development, Shandong University, Weihai 264209, China

⁶ The Bartlett School of Sustainable Construction, University College London, London WC1E 6BT, UK

⁷ MOE Engineering Center of Forestry Biomass Materials and Bioenergy, Beijing Forestry University, Beijing 100083, China

⁸ School of Public Policy and Administration, Xi'an Jiaotong University, Xi'an 710049, China

⁹ Senckenberg Biodiversity and Climate Research Centre (SBIK-F), Frankfurt am Main D-60325, Germany

¹⁰ Lead contact

1

2 * Correspondence: chen_bin@fudan.edu.cn (B.C.), gqchen@pku.edu.cn (G.C.)

1 **Summary:** The continued loss of unfragmented intact forest landscapes (IFLs)
2 despite numerous global conservation initiatives indicates the need for improved
3 knowledge of proximate and underlying drivers. Yet the role of non-agricultural
4 activities in forest degradation and fragmentation has not received adequate attention.
5 We focus on IFL loss caused by various economic activities and investigate the
6 influence of global consumption and trade via the multi-regional input-output model.
7 For IFL loss associated with the 2014 world economy, over 60% was related to final
8 consumption of non-agricultural products. More than one-third of IFL loss was linked
9 to export, primarily from Russia, Canada and tropical regions to mainland China, the
10 EU and the USA. Of IFL loss associated with export, 51% and 26% was directly caused
11 by logging and mining/energy extraction, respectively. The dispersed nature of IFL loss
12 drivers and their indirect links to individual final consumers call for stronger
13 government engagement and supply-chain interventions.

1 **Introduction**

2 Intact forest landscapes (IFLs), defined as continuous expanses of forests and
3 associated ecosystems detecting no evidence of human interference and having
4 sufficient space to maintain native biodiversity (a minimum size of 500 km²)¹, play an
5 irreplaceable role in delivering ecosystem services, such as regulating climate^{2,3} and
6 harboring biodiversity^{4,5}. For example, IFLs have higher resilience to natural
7 disturbance (e.g., climate change) and store much more carbon per hectare than other
8 forest zones (over 3 times higher in Latin America and Africa)¹. Boreal IFLs also
9 contain a large fraction of Arctic permafrost, which is a crucial organic carbon reservoir
10⁶. Possibly due to insufficient knowledge of both proximate and underlying drivers of
11 IFL loss, global IFL area reduction has reached 1.5 million km² during the last two
12 decades, more than quadruple the area of Germany. Only 20% of the global tree cover
13 was within IFLs by 2020⁷. Carbon emissions from IFL loss may compromise global
14 net-zero strategies and hinder the implementation of land-based climate solutions.

15 There is widespread concern about the driving effects of local production on
16 deforestation. Agricultural expansion as the primary proximate driver of deforestation
17 has been well acknowledged and broadly studied⁸. However, regional land use change
18 is no longer simply driven by local demand, but is rather indirectly influenced by
19 international markets and the surging consumption of land-based products⁹. Countries
20 with forest conservation goals can import finished land-based products via global
21 supply chains, displacing land-use pressure and related eco-environmental impacts
22 outside their own territory borders. For example, Russia produces large amounts of
23 wood for forest-scarce or strictly regulated regions (e.g., China and the EU), which puts
24 pressures on Russia's domestic IFLs¹⁰. Such globalization of land use has spurred
25 substantial studies on consumption-side drivers of land use^{11,12}, deforestation^{13,14} and
26 corresponding environmental impacts^{15,16}, but none have focused specifically on IFL
27 loss.

28 In addition to deforestation (i.e., the complete removal of tree cover followed by

1 a change in land use), forest ecosystems are also facing emerging threats from
2 degradation and fragmentation. Even the removal of narrow tracts of forests can affect
3 overall forest structure and composition, inducing landscape-scale ecological changes
4 (e.g., tree mortality and biodiversity loss) ^{17,18}, increasing vulnerability to external
5 disruption (e.g., fires and wind) and initiating a cascade of land use changes because of
6 easier access to the remaining forests ^{19,20}. Forest degradation and fragmentation caused
7 by logging sites and roads often precede deforestation ²¹. In particular, mining and
8 energy extraction has become a dominant culprit second only to agriculture, due to
9 fragmenting effect of narrow exploration trails and electricity transmission lines ²².
10 Carbon loss resulting from forest degradation has also exceeded that from deforestation
11 in the Brazilian Amazon ²³.

12 Since the framework of IFL mapping emphasizes the critical role of forest
13 intactness and size, degradation and fragmentation from forestry, mining and energy
14 extraction can receive due attention. For example, an 800 km² of primary forest that is
15 bisected by a road into two tracts less than 500 km² each would be identified as 800
16 km² of IFL loss. There can also be IFL loss with no or little deforestation, if primary
17 forests are replaced by planted forests. As the trade of industrial roundwood, fossil fuels,
18 metal and minerals represents a large portion of global total production ²⁴⁻²⁶, tele-
19 connecting regional IFL loss to distant consumption can provide a novel perspective of
20 how global supply chains of various non-agricultural commodities influence worldwide
21 forest ecosystems. Considering the exceptional conservation value of IFLs in terms of
22 stabilizing terrestrial carbon stocks and harboring biodiversity, IFL loss displacement
23 can also reflect potential indirect driving forces behind carbon emissions and
24 biodiversity loss.

25 Given all this, we provide a comprehensive overview of IFL loss associated with
26 global supply chains, with a specific focus on the distinct roles of various economic
27 activities as well as their links to domestic consumption and export production. To do
28 so, we integrate a spatially-explicit global dataset on IFL loss with the multi-regional

1 input-output model based on the Global Trade Analysis Project (GTAP) database ²⁷.
2 Our results show that, for IFL loss associated with the 2014 world economy, 37% was
3 related to export production destined for global markets, especially mainland China, the
4 EU and the USA, of which over three quarters was directly caused by logging, mining
5 and energy extraction. More than 60% of the overall IFL loss was linked to the final
6 consumption of a highly dispersed range of non-agricultural commodities, with indirect
7 links to individual final consumers. Therefore, distinct from agriculture-dominated
8 deforestation, tackling IFL loss requires stronger engagement of national governments
9 complemented with supply-chain and demand-side interventions.

10

11 **Results**

12 **IFL loss embodied in final consumption**

13 The global dataset classified IFL loss into 5 proximate causes, including wildfires,
14 industrial logging, agricultural expansion, mining/oil and gas/hydropower as well as
15 transportation for other aims, and we only considered IFL loss caused by human
16 economic activities (excluding wildfire). Of all the anthropogenic IFL loss, industrial
17 logging, agricultural expansion and mining/energy contributed 47%, 35% and 15%,
18 respectively. IFL loss is accordingly allocated to the primary economic sectors directly
19 responsible for the loss and then traced through downstream intermediate producing
20 sectors (which use primary products as intermediate input) eventually to the final
21 consumption of IFL-risk final products.

22 Fig. 1 presents the IFL loss associated with the final consumption of six region
23 groups (consistent with World Bank definitions) and dominant final consumers. Most
24 of major producer and consumer countries are listed in the GTAP database, but the
25 Democratic Republic of the Congo, Gabon and Republic of the Congo, which are
26 among the top 10 countries with the most widespread IFL loss, are aggregated to South
27 Central Africa or Rest of Central Africa. At a national scale, the attribution of embodied
28 IFL loss was dominated by hotspots of IFL loss (see Fig. 2) as well as large consumer

1 countries. Brazil was the largest final consumer of IFL-risk products (related to 0.92
2 Mha of global IFL loss, 16%), followed by Russia (0.59 Mha, 11%), mainland China
3 (0.57 Mha, 10%), the USA (0.31 Mha, 6%), Canada (0.28 Mha, 5%), Bolivia (0.27 Mha,
4 5%), Peru (0.24 Mha, 4%), Indonesia (0.23 Mha, 4%) and South Central Africa (0.20
5 Mha, 4%). The embodied IFL loss of other countries and regions was less than 0.20
6 Mha. For the six region groups, Latin America accounted for the largest share of
7 embodied IFL loss (35%), mainly due to extensive IFL area reduction locally. Asia-
8 Pacific and Europe & Central Asia accounted for approximately 24% and 19%
9 respectively, as a combined result of the high consumption level of developed and
10 emerging countries and large-scale IFL loss in a handful of countries within these
11 regions (e.g., Indonesia and Russia). Only 10% of global IFL loss was associated with
12 the final consumption of Sub-Saharan Africa.

13 When linking proximate causes of IFL loss to final consumption sectors (Fig. 1D),
14 it can be observed that, IFL loss from agricultural expansion was primarily associated
15 with the final consumption of food, but considerable logging- and mining/energy-
16 driven IFL loss was embodied in construction, tertiary and other secondary sectors. This
17 means that some of the primary products from logging and mining/energy (e.g.,
18 roundwood and crude oil) were used by these sectors as intermediate inputs to produce
19 final products and support services. Overall, only 39% of IFL loss was linked to
20 agriculture-related products (including primary and processed food, textiles and
21 wearing apparel), while 18% was linked to forest, metal, mineral and energy products
22 and the rest to services and highly processed products.

23 The final consumption sectors associated with IFL loss varied greatly across
24 countries and regions. In Latin America, 68% of IFL loss was embodied in the final
25 consumption of agriculture-related products, while raw forest products, timber and
26 paper took the lead in South Central Africa (56%). A notable share of IFL loss was
27 embodied in construction in Indonesia (49%), mainland China (45%), Canada (36%),
28 Cameroon (31%), Japan (27%), India (25%) and Russia (21%), and final consumption

1 of mineral, metal and energy products represented a relatively high proportion of IFL
2 loss in Russia (14%). There was also a large amount of IFL loss embodied in other
3 secondary sectors and tertiary sectors, especially in mainland China (36%) and
4 developed countries (e.g., the USA, Canada, Japan and Germany).

5

6 **IFL loss embodied in international trade**

7 Fig. 2A explores the linkages between producers of primary IFL-risk products and
8 final consumers of final products. IFL loss associated with distant final consumption
9 comprised 37% of global total anthropogenic IFL loss. Generally, displaced IFL loss
10 was linked to the exports from IFL-rich regions (e.g., Russia, Canada, Southeast Asia,
11 Oceania, Central Africa and Latin America) to developed (i.e., the EU and the USA)
12 and emerging economies (i.e., mainland China). Meanwhile, source regions of
13 mainland China's imports were more diversified than those of the EU and the USA.

14 Specifically, exports from Brazil and Rest of Latin America to the EU, mainland
15 China and the USA were related to 103 kha, 85 kha and 81 kha of IFL loss respectively.
16 The relatively small impact of mainland China on IFLs in Rest of Latin America is
17 because pasture expansion comprised over 80% of the total agriculture-related IFL loss
18 in tropical Latin America, while according to the Food and Agriculture Organization
19 (FAO) FAOSTAT production and trade data ²⁸, mainland China mainly imported cattle-
20 related products from countries with little IFL loss, such as Argentina and Uruguay. For
21 Russia and Africa, the major importers of IFL-risk commodities were the EU and
22 mainland China, but the EU relied more on Russia (138 kha) and mainland China
23 imported more from Africa (137 kha). For Canada, the USA was the leading export
24 market, responsible for 139 kha of IFL loss, 100 kha more than that exported to
25 mainland China. IFL-risk products originating from Oceania and Southeast Asia were
26 predominantly sold to mainland China, linked to 146 kha and 73 kha of IFL loss,
27 respectively.

28 Considering the heterogeneities of IFLs across different climate zones in terms of

1 their potential to sequester carbon, conserve biodiversity and deliver other ecosystem
2 services, Fig. 2B further classifies IFL loss into different biomes based on the spatial
3 dataset “Terrestrial Ecoregions of the World” provided by the World Wildlife Fund ²⁹,
4 and investigates their connection with final consumers. Seventy percent of the global
5 anthropogenic IFL area reduction took place in tropical and subtropical biomes
6 (basically in Latin America, Central Africa, Southeast Asia and Oceania), while boreal
7 (mostly in Russia and Canada), temperate (mostly in Canada) and Mediterranean
8 (mostly in Australia) biomes accounted for 22%, 5% and 2%, respectively. For final
9 consumers, the structure of source biomes varied between countries/regions. Of all the
10 IFL loss embodied in mainland China’s final consumption, over 2/3 was sourced from
11 tropical and subtropical biomes and 15%, 8% and 7% came from boreal, temperate and
12 Mediterranean biomes, respectively. The EU and the USA imported less share of IFL-
13 risk products from tropical and subtropical biomes (52% and 42%, respectively),
14 whereas boreal regions contributed approximately 40% for each of the two. For major
15 final consumers who were also dominant primary producers, final demand was mainly
16 satisfied at the expense of local IFL loss. Other regions’ final consumption generally
17 threatened more tropical and subtropical IFLs than IFLs in other biomes. Overall, 1/3
18 of tropical and subtropical IFL loss was driven by export production, but the ratio was
19 over 40% for boreal and temperate IFL loss.

20

21 **Displaced IFL loss by proximate causes**

22 When tracing embodied IFL loss back to its proximate causes, it can be found that,
23 different countries and regions influenced IFLs through distinct pathways (Fig. 3).
24 Logging-driven IFL loss made up the largest share of total displaced IFL loss (1.04 Mha,
25 51%), followed by mining and energy extraction (0.53 Mha, 26%) and agriculture (0.46
26 Mha, 22%). Logging-driven IFL loss was primarily embodied in the exports from
27 Canada, Russia, Southeast Asia, Oceania and Sub-Saharan Africa to East Asia
28 (especially mainland China), the USA and the EU. IFL loss caused by other economic

1 activities was associated with similar importers but can be traced back to different
2 source regions. 83% of displaced agriculture-related IFL loss originated from Latin
3 America, with Brazil alone contributing 42%, while displaced IFL loss from mining
4 and energy extraction primarily came from Russia (43%), Australia (22%) and Latin
5 America (28%). Generally, 62% of mining/energy-induced IFL loss was associated
6 with export production, but the shares for logging and agriculture only reached 40%
7 and 23% respectively. In regard to final consumption, logging-related IFL loss was
8 linked to consumers worldwide, led by China as an individual economy. Almost 70%
9 of the agriculture-related IFL area reduction was associated with the final consumption
10 of Latin America, while Russia and the EU were responsible for the largest share of IFL
11 loss from mining and energy extraction (23% and 17% respectively).

12 Fig. 4 further illustrates the displacement of IFL loss by proximate causes for
13 major primary producers and final consumers. For Latin America, Central Africa,
14 Canada, Russia and Indonesia, approximately 23% – 49% of local IFL loss was related
15 to export, but the ratio reached 78% and 65% for Australia and Rest of Oceania &
16 Southeast Asia. There are obvious differences in the share exported by different
17 proximate causes for different source regions. In Latin America, IFL loss from
18 agriculture was exported to a much larger degree than that from mining/energy and
19 logging, but local demand still represented 76% and 83% of the total agriculture-
20 induced IFL loss in Brazil and Rest of Latin America, respectively, which is consistent
21 with FAO production and trade data. For Central Africa, Canada, Indonesia and Rest of
22 Oceania & Southeast Asia, most of the IFL loss embodied in export was sourced from
23 logging. The contribution of external demand to IFL loss in Central Africa was lower
24 than potentially expected. This lower damage is because even though forestry
25 companies in the DR Congo are typically foreign-owned, a large portion of their output
26 is aimed for domestic markets³⁰; meanwhile, for Gabon and Republic of Congo where
27 natural forests are publicly owned, FAO Yearbook of Forest Products also confirms that
28 domestic demand were responsible for a substantial share of wood-based products³¹.

1 Mining and energy extraction was the dominant proximate driver of IFL loss embodied
2 in exports from Russia and Australia, even though logging-related IFL loss was more
3 pervasive in Russia. The import structures by proximate causes for China and the USA
4 were similar: over 60% of the IFL loss embodied in their imports came from logging,
5 followed by mining/energy (contributing 19% and 25%, respectively). For the EU,
6 mining/energy (41%) prevailed over the other two proximate causes.

7

8 **Discussion**

9 We have comprehensively investigated IFL loss across global supply chains and
10 revealed distinctly different patterns that complement studies solely focusing on
11 deforestation. There are some uncertainties in our estimation because multiple data
12 sources and certain attribution models are adopted to allocate IFL loss to different
13 economic sectors and to different years due to a lack of detailed data. Sectoral and
14 regional aggregations in the multi-regional input-output model also preclude us from
15 identifying supply chain agents at a higher resolution and may lead to the misestimation
16 of the influence of international trade. However, both qualitative insights from the
17 literature and quantitative insights from sensitivity analyses support the reliability of
18 our findings. Please see the Uncertainties and Limitations section for more details.

19 Our results uncovered the notable contribution of non-agricultural products in
20 shaping IFL loss displacement: logging-induced IFL loss comprised the largest share;
21 More strikingly, although the IFL loss from agriculture was more than double that from
22 mining and energy extraction, the displaced IFL loss that was directly caused by mining
23 and energy extraction was larger in magnitude. These results highlight the need to
24 assess the broader impacts of global supply chains, not only on agriculture-dominated
25 deforestation, but also on the prevalent forest fragmentation and degradation from
26 various economic activities. However, selective logging is still promoted as a
27 sustainable management practice by many tropical countries and accepted by some
28 influential international programs (e.g., Reducing Emissions from Deforestation and
29 Forest Degradation (REDD) program and Forest Stewardship Council (FSC)

1 certification program) ³², even though it is becoming an important agent of
2 fragmentation and degradation of the world's most precious tropical forests ^{18,33}.
3 Current pledges made by companies also pertain to deforestation from traditional forest
4 and agricultural commodities, such as timber, pulp, soy, beef and palm ³⁴, neglecting
5 the threats from mining and energy extraction, which will be more pivotal in the
6 ongoing global transition towards a green energy system ³⁵. Moreover, a large share of
7 IFL loss was embodied in downstream sectors, such as construction, service and
8 manufacturing of highly processed products. Trade analyses solely focusing on the
9 primary products directly responsible for forest disturbance can greatly underestimate
10 IFL loss displacement and consequently mislead demand-side policies. This reflects the
11 relevance of land use analyses from a consumption-based perspective.

12 Given the pervasive displacement of IFL loss, mutual efforts of producer and
13 consumer countries are required to preserve remaining IFLs. There have already been
14 many voluntary initiatives against deforestation. Examples on the supply side include
15 corporate pledges (e.g., zero-deforestation commitments) and collective aspirations
16 (e.g., New York Declaration on Forests). Targeting IFL loss other than deforestation can
17 complement existing efforts in these aspects. For instance, among all the companies
18 with a zero-deforestation commitment, 44% adopt a net-zero target, which allows
19 afforestation to compensate deforestation of the same size ³⁴. This may lead to the
20 situation where primary forests rich in biodiversity and carbon stocks are replaced by
21 managed forests. Such conversion could be mitigated if a zero-IFL-loss target is
22 integrated into the commitment, as the concept distinguishes managed and natural intact
23 forests. There are also various consumer campaigns and supply chain initiatives, such
24 as the Roundtable for Sustainable Palm Oil, the Roundtable for Responsible Soy and
25 the Programme for the Endorsement of Forest Certification. FSC has introduced rules
26 for IFL protection since 2017, but it allows 20% of IFLs to be exploited ³⁶. Apart from
27 the problem of leniency, this type of approach will also be hard to duplicate for many
28 other IFL-risk commodities, because their links to IFL loss are more indirect for

1 individual final consumers compared to traditional agricultural and forest products. For
2 example, it is widely concerned that beef production drives deforestation in the Amazon,
3 but it is hard for consumers to realize that the production of highly processed equipment
4 may involve timber and metals produced at the expense of IFL loss and that services
5 provided by tertiary sectors may be supported by electricity generated from oil and gas
6 associated with IFL loss. Voluntary measures may also suffer from other problems such
7 as low/selective adoption, insufficient market uptake, corruption and patrimonialism ³⁴.

8 The much more dispersed nature of IFL loss drivers and the weaknesses of private
9 interventions call for stronger engagements of both national governments and
10 international institutions. For instance, producer governments can introduce mandatory
11 due diligence for corporations and transparency regulations for financial institutions ³⁷.
12 It is also crucial to implement more stringent land use policies, e.g., through land-use
13 zoning policies, such as protected areas and biodiversity corridors. Variations in IFL
14 loss drivers across countries and sectors require tailored forest conservation strategies.
15 For Latin America, combating agricultural encroachment is the primary challenge. For
16 Russia, Canada, Southeast Asia, Oceania and Central Africa, policies should target
17 industrial logging, with mining/energy extraction as another focus for Russia and
18 Canada and agricultural expansion for other regions. Meanwhile, consumer countries
19 and international institutions can support producer countries (especially where forest
20 conservation institutions are lacking or severely underfunded) in specific initiatives or
21 general capacity building, such as improving governance, ensuring land rights,
22 enhancing productivity and establishing a monitoring and verification system that can
23 improve the traceability and transparency of supply chains. International demand may
24 also spark illegal forest development: approximately 50% and 25% of illegal timber in
25 international trade came from Indonesia and Brazil in 2013, respectively ³⁸. Consumer
26 countries can therefore enforce regulations to combat illegal logging and deforestation
27 (e.g., EU Forest Law Enforcement, Governance and Trade (FLEGT) Action Plan), for
28 example, by validating the source of imported products. Of course, there are many other

1 policy options that have not been listed here, such as “nudging” and behavioral designs,
2 carbon tax, ecological payments and trade tariffs. However, considering the complex
3 trade-offs between potential outcomes and policy feasibility, all of the proposed options
4 should be explored and adopted with caution ³⁷.

5 In the future, forests are facing fiercer threats from multiple sectors, as more
6 resources are needed to feed a growing population and to support quality life for all. In
7 this context, advancing the knowledge of IFL loss has become increasingly crucial.
8 However, neither the proximate IFL loss cause nor its connection with complex socio-
9 economic dynamics has been sufficiently studied. The problem, to a large extent, is due
10 to a lack of data suitable for comprehensive and in-depth interdisciplinary analyses.
11 Therefore, researchers from land-use, geographic and social sciences should strengthen
12 cooperation, to develop a robust globally consistent dataset that has higher spatial,
13 temporal and commodity resolution. Beyond risks from different sectors, there are also
14 trade-offs between different environmental and socio-economic targets. For example,
15 commercial logging and subsistence agriculture in forest areas can help eradicate
16 poverty and underpin local livelihood ³⁹. Smallholder subsistence agriculture is an
17 important source of employment and livelihoods for women in rural places (supporting
18 more than two-thirds of working women in Africa) ⁴⁰. At least 36% of the world’s IFLs
19 are within indigenous and tribal territories ⁴¹. Unbalanced conservation policies may
20 impede social stability and the achievement of United Nations Sustainable
21 Development Goals (SDG), such as SDG No.1 (No poverty), SDG No.2 (Zero hunger)
22 and SDG No.5 (Gender equality). As a result, it is critical that the synergies and trade-
23 offs between sustainable development goals in different domains are fully understood
24 and considered in policy-making. The coordination efforts require multidisciplinary
25 planning as well as voices from different interest groups across different scales.

26

1 **Experimental Procedures**

2 **Resource Availability**

3 Lead contact: Further information and requests should be directed to and will be
4 fulfilled by the lead contact, Bin Chen (chen_bin@fudan.edu.cn).

5 Materials availability: This study did not generate new unique materials.

6 Data and code availability: All the data sources and mathematic models used in this
7 paper are elaborated in the Experimental Procedures and summarized here, including
8 loss of intact forest landscapes ¹, drivers of global forest loss ³³, global mining areas ⁴²,
9 FAO land use ⁴³, principal mineral areas, producing mines, and oil and gas fields in
10 Canada ⁴⁴, locations of global hydropower plants ⁴⁵, the European Space Agency (ESA)
11 CCI global land cover maps ⁴⁶ and GTAP multi-regional input-output table ²⁷. Data on
12 the loss of intact forest landscapes associated with domestic consumption, import and
13 export are available in the Supplemental Information. Any additional information
14 required to reanalyze the data reported in this paper will be shared by the lead contact
15 upon request.

17 **Attribution of IFL loss to economic sectors**

18 Data on IFL loss were taken from Potapov et al.¹, which recorded IFL area
19 reduction per country in 2000-2013, based on the global archive of Landsat satellite
20 imagery supplemented with national transportation maps, high-resolution maps from
21 Google Earth and other forest cover change products. They then applied a stratified
22 sampling approach to identify the proximate causes of IFL loss for 9 aggregated regions,
23 which were further examined with all accessible remote sensing data. The proximate
24 drivers were divided into five categories, including wildfires, industrial logging,
25 agricultural expansion (including pasture expansion), mining/oil and gas/hydropower
26 as well as transportation for other aims. IFL loss from wildfires was not included in our
27 calculation.

28 We first need to identify IFL loss by proximate drivers per country. By integrating
29 the geographic boundaries of the 9 aggregated regions and the data on IFL loss, we can
30 obtain driver-specific IFL loss for Canada, Russia and Australia. For other countries,
31 we took the following procedures: (1) the IFL loss map was overlapped with a map of
32 agriculture-driven forest cover loss ³³, in order to estimate each country's relative
33 contribution to regional total agriculture-induced IFL loss. Then, the IFL loss from

1 agriculture per country was calculated by multiplying the regional total by the
2 corresponding shares. (2) IFL loss from mining and energy extraction per country was
3 calculated following the same principle in (1). To assess each country's share of
4 regional total mining/energy-related IFL loss, the IFL loss map was overlapped with a
5 map of global mining areas ⁴². We applied a 70 km buffer around mining sites, given
6 that forest disturbance may extend up to 70 km from mining leases ⁴⁷. (3) Regarding
7 IFL loss from other transportation, which represented only 2% of global total
8 anthropogenic IFL loss, we assumed it was proportional to national overall IFL loss. (4)
9 Finally, logging-related IFL loss per country was obtained by subtracting the IFL loss
10 caused by other proximate drivers from the national total. This is because a large
11 proportion of logging-related IFL loss was due to degradation and fragmentation from
12 selective logging, in which case the disturbed landscapes remained forest landscapes in
13 land cover maps and were more difficult to distinguish from forest wildlands compared
14 to agriculture and mining areas.

15 We then allocated IFL loss to specific economic sectors in the GTAP input-output
16 table (IOT) as a base for multi-regional input-output analysis, in order to track IFL loss
17 across global supply chains. IFL loss from logging and other transportation can be
18 directly attributed to the forestry and land/pipeline transport sectors respectively, but
19 IFL loss due to other causes needs to be disaggregated to fit IOT sectoral resolution.
20 Following Pendrill et al. ^{14,16}, areas occupied for different agricultural uses were
21 attributed in relative proportion to the expansion of each agricultural sector's direct land
22 use, based on the land use data from FAO ⁴³. For example, if wheat represented x% of
23 the total agriculture expansion, x% of the agriculture-related IFL loss would be assigned
24 to the wheat production sector. Potapov et al. noted that pasture expansion contributed
25 81.5% of the overall agriculture-related IFL loss in tropical South America, so 81.5%
26 of IFL loss was attributed to cattle farming before we estimated the contribution of other
27 agricultural sectors. Most of mining/energy-related IFL loss took place in tropical South
28 America, Australia, Canada and Russia. For Latin America and Australia, mining

1 (mostly gold exploration) was the primary cause, therefore, IFL loss was attributed to
2 the mining sector ¹. For Canada, we separated the contributions of mining, oil and gas
3 extraction and hydropower production by integrating the IFL loss map, the
4 abovementioned global mining map ⁴², the map of oil and gas field provided by Natural
5 Resources Canada ⁴⁴ and the geographic coordinates of global hydropower plants
6 provided by Global Power Plant Database ⁴⁵. Sonter et al. found that off-lease impacts
7 of mining within surrounding 70 km buffers were due to secondary forest clearing,
8 urban expansion to support the workforce and other mining-stimulated economic
9 activities ⁴⁷. Such a cascade influence should also exist during oil, gas and hydropower
10 production, so we consistently assume a 70 km impact buffer for all these sites. In
11 Russia, oil and gas extraction was the disturbing factor ¹, where IFL reduction was
12 equally assigned to the oil and gas sectors due to a lack of data.

13

14 **Amortization of IFL loss**

15 Attributing IFL loss to drivers implies not only specifying which activities (i.e.,
16 economic sector or production of a given commodity) cause the loss, but also
17 accounting for the temporal dimension of the link between IFL loss and economic
18 activities. That is, while IFL loss is a one-off event, the follow-up production typically
19 persists over long timeframes. For example, an oil palm or acacia plantation or a bauxite
20 mine established within an intact forest area will generate palm oil, pulp and aluminum
21 products over many years, not seldom with a time-lag between IFL loss and production
22 (oil palms only start generating fruit after three years, and acacia pulp plantations
23 typically have a rotation period of seven years) ^{48,49}. For this reason, one cannot simply
24 assign IFL loss to economic activities in the same year, but land-use changes (as well
25 as associated environmental impacts) are typically amortized, or spread out, over
26 several years of production. This means that with an amortization time of T years, the
27 IFL loss embodied in economic activities in a given year should be the total IFL loss in
28 the previous T years divided by T.

1 The choice of amortization period is ultimately normative and there is no agreed-
2 upon choice. In life cycle assessment, an amortization period of 20 years is typically
3 advocated ⁵⁰, but different amortization periods are used in the literature on
4 deforestation embodied in trade (typically in the range of 1-10 years) ^{48,49}. Here we
5 adopted an amortization period of 13 years for pragmatic reasons: the IFL data are
6 aggregated over the period 2000-2013 and cannot easily be disaggregated to an annual
7 time-series (i.e., there are no obvious proxies that can be used to infer annual IFL loss
8 in the period). Unless there are large fluctuations in IFL loss in a given country or region,
9 the choice of a thirteen-year amortization period should not affect our estimation of the
10 roles of different countries in global supply chains ⁴⁸.

11 It is common that forestry land is later converted to agricultural land ⁵¹, and part
12 of the IFL loss should be allocated to the follow-up activities. According to Potapov et
13 al., in Africa and Southeast Asia, logging-induced IFL loss was caused by selective
14 logging, which means that the landscape remained forest land cover and would not be
15 easily occupied by agriculture and other industries. Monoculture plantations were
16 found to follow selective logging but only contributed 0.2% of the global total IFL area
17 reduction. Considering that it is also unknown since when logging sites were converted
18 to plantations, IFL loss was not attributed to follow-up production. In Latin America,
19 there is no specific description about which kind of logging method was used, but it is
20 stated that new cropland mainly occurred in pastures previously converted from forests,
21 which is also confirmed by ref. ⁵¹. To estimate the successive land use change, we
22 overlapped the 300-m-resolution ESA-CCI global land cover maps ⁴⁶ with the IFL loss
23 map. The class “Grassland” in ESA CCI maps was used to represent pastures (existing
24 time-series high-resolution global land cover maps do not distinguish natural grassland
25 and pastures) and the class “Agriculture” (including grids of cropland and grids of
26 mosaic cropland/natural vegetation) was used to represent cropland, in order to estimate
27 the maximum probability of pasture-cropland conversion on an annual basis. No grids
28 detecting pasture-cropland conversion in the period concerned were found to overlap

1 with IFL loss patches. In North America and Eurasia, clear-cutting was the main culprit
 2 of logging-related IFL loss. Meanwhile, mining/energy-related IFL loss was associated
 3 with oil and gas production in Eurasia and primarily associated with hydropower
 4 production in North America, and we inferred that these oil, gas and hydro fields would
 5 not be used for other aims during such a short period. We followed the same method
 6 mentioned above to estimate the maximum probability of agricultural encroachment
 7 into logging and mining areas. Little grids detecting new agricultural land since 2000
 8 were within lost IFLs, even though we have overestimated the potential distribution of
 9 agricultural land.

10

11 **Embodiment accounting**

12 The embodiment accounting is carried out based on multi-regional input-output
 13 (MRIO) analysis, which has been adopted to account for a variety of ecological
 14 elements (e.g., energy ^{52,53} and carbon emissions ^{54,55}). Within the framework of the
 15 multi-regional input-output model, the world economic system consists of m regions,
 16 each with n sectors. Intermediate trade denotes transactions between industrial sectors
 17 while final trade depicts transactions between sectors and final consumers. Land use
 18 responsible for IFL loss is embodied in IFL-risk products, and therefore can be traced
 19 as product embodiments from countries witnessing IFL loss to final consumers of
 20 related products. Sectoral input of IFL-risk land use comprises IFL loss induced by the
 21 sector directly (marked as l), and that embodied in imported intermediate products (z),
 22 while sectoral output consists of IFL loss embodied in all the outputted products in both
 23 intermediate and final trade (f). As elaborated by the law of conservation of resource
 24 use, total sectoral input equals sectoral output ⁵⁶, which generates the following
 25 equation for sector i in region r :

$$l_i^r + \sum_{s=1}^m \sum_{j=1}^n (\varepsilon_j^s z_{ji}^{sr}) = \varepsilon_i^r \left(\sum_{s=1}^m \sum_{j=1}^n z_{ij}^{rs} + \sum_{s=1}^m f_{ic}^{rs} \right) \quad (1)$$

26 where z_{ji}^{sr} stands for intermediate input to sector i in region r from sector j in

1 region s . f_{ic}^{rs} represents the final output from sector i in region r to region s for final
 2 consumption. ε_i^r denotes the IFL loss associated with the unit output of sector i in
 3 region r .

4 Equation (1) can be transformed into matrix form:

$$\mathbf{L} + \boldsymbol{\varepsilon}\mathbf{Z} = \boldsymbol{\varepsilon}\mathbf{X} \quad (2)$$

5 where $\boldsymbol{\varepsilon} = (\varepsilon_i^r)_{1 \times mn}$, $\mathbf{L} = (l_i^r)_{1 \times mn}$, $\mathbf{Z} = (z_{ij}^{rs})_{mn \times mn}$, $\mathbf{X} = \text{diag}(x_i^r)_{mn \times mn}$
 6 ($x_i^r = \sum_{s=1}^m \sum_{j=1}^n z_{ij}^{rs} + \sum_{s=1}^m f_{ic}^{rs}$, denotes the sectoral total output). Therefore, $\boldsymbol{\varepsilon}$ can be
 7 obtained as:

$$\boldsymbol{\varepsilon} = \mathbf{L}(\mathbf{X} - \mathbf{Z})^{-1} = \mathbf{L}\mathbf{X}^{-1}(\mathbf{I} - \mathbf{A})^{-1} \quad (3)$$

8 where \mathbf{I} is the identity matrix with dimensions $mn \times mn$ and $\mathbf{A} = \mathbf{Z}\mathbf{X}^{-1}$.

9 IFL loss embodied in the sectoral input/output can be obtained by multiplying their
 10 volume by the corresponding intensity ε_i^r . For example, for region r , the IFL loss
 11 embodied in its final consumption (LC^r) can be calculated as:

$$LC^r = \sum_{s=1}^m \sum_{j=1}^n \varepsilon_j^s f_{jc}^{sr} \quad (4)$$

12 Multi-regional input-output tables (IOTs) were collected from the GTAP database
 13 ²⁷, given that GTAP covers 141 countries, which enables investigation of detailed
 14 information for major producer and consumer countries and regions, and 65 sectors,
 15 which allows us to distinguish the impacts of different economic sectors.

16

17 **Uncertainties and Limitations**

18 First, there are some uncertainties in the attribution of IFL loss. The criteria and
 19 method to identify IFL loss and its primary causes are more complicated than those to
 20 identify deforestation. The massive remotely sensed data as well as data processing
 21 required make it even more challenging to replicate the work of Potapov et al., not to
 22 mention allocate IFL loss to economic sectors at higher sectoral resolution, for lack of
 23 high-resolution spatial information on land use. Therefore, we relied on multiple data

1 sources and certain attribution model to estimate IFL loss by proximate drivers and by
2 specific sectors at a national scale. The model we applied and the inconsistencies
3 between different data sources (e.g., spatial resolution and identification approach) may
4 cause some deviations. Therefore, we provided a sensitivity analysis (abbreviated
5 below as SA for convenience of presentation) by assuming that the IFL loss caused by
6 each proximate driver fluctuates around the baseline (values obtained based on our
7 method) by 10%, 20% and 30% of each country's total IFL loss respectively (SA-
8 driver), and assuming that the IFL loss allocated to each sector fluctuates around the
9 baseline by 10%, 20% and 30% of each country's IFL loss from the corresponding
10 driver respectively (SA-sector), on the premise that all the results are consistent with
11 the findings of Potapov et al (e.g., total IFL loss per country and driver-specific IFL loss
12 per region). As mentioned above, we can obtain driver-specific IFL loss for Canada,
13 Russia and Australia from existing data, so the SA-driver analyses were only performed
14 for Latin America, Asia-Pacific (excluding Australia) and Africa (see Supplemental
15 Information). The results show that different allocation schemes mainly affect IFL loss
16 associated with the final consumption of producer countries, mainland China and the
17 USA, because IFL loss in producing countries would be allocated to sectors that were
18 more/less export-oriented in SA, while mainland China and the USA were dominant
19 export markets. Such influence is very limited in the SA for the Asia-Pacific, Africa,
20 Russia and Canada while it is more evident in the SA for Latin America, but generally,
21 dominant final consumers still hold their leading positions.

22 Second, the MRIO model has some uncertainties. The input and output of
23 individual companies are aggregated to a sectoral level in input-output tables, and it is
24 also unknown which companies were responsible for the IFL loss, precluding us from
25 identifying whether IFL-risk commodities were sold to domestic or foreign markets
26 specifically. However, local IFL loss is a combined result of both internal and external
27 demand. For example, if there is no timber consumption domestically, then foreign
28 demand can be satisfied by timber production in existing logging sites, avoiding further

1 expansion of forestry land use into IFLs. In this sense, it is reasonable to use aggregate
2 sector-level trade data rather than company-level trade data to reveal the potential and
3 indirect driving effects of distant consumers. Meanwhile, the DR Congo is part of the
4 aggregated region South Central Africa and Gabon and Congo are part of the aggregated
5 region Rest of Central Africa. Production and trade data on roundwood, sawnwood,
6 wood-based panels and wood charcoal show that, the share of export in total production
7 for South Central Africa and for the DR Congo were very similar, while the share for
8 Rest of Central Africa was slightly lower than the share for Congo and Gabon ³¹.
9 Therefore, the IFL loss associated with exports from Rest of Central Africa might be
10 underestimated. The abovementioned uncertainties from sector and spatial aggregation
11 have been well addressed by previous studies ⁵⁷. Physical accounting is another
12 frequently applied tool, which utilizes physical bilateral trade data on primary and
13 processed commodities and therefore can trace resource use and environmental impacts
14 at a more detailed commodity level ⁵⁸. However, this advantage is diminished here,
15 because detailed information on responsible companies/commodities is lacking. In
16 addition, physical accounting has limitations in capturing flows of highly processed
17 products (to which a large share of IFL loss was linked) and determining actual end
18 users.

19 Third, there are also uncertainties arising from the amortization of IFL loss. The
20 length of the amortization period may affect our results, but we were not able to perform
21 sensitivity analyses because there are no appropriate references to disaggregate IFL loss
22 to an annual time-series and the recent data on IFL loss in 2014-2020 do not distinguish
23 specific proximate drivers ⁷. Meanwhile, considering trade-offs between time, country
24 and sector resolutions, we adopted the 2014 GTAP IOT for MRIO analysis (there is no
25 IOT for 2013), assuming that the difference between the IFL loss in 2000-2013
26 amortized over 13 years and the IFL loss in 2000-2014 amortized over 14 years can be
27 ignored. According to Global Forest Watch ⁵⁹, loss of global primary forests (here $\geq 20\%$
28 canopy density) did not accelerate until 2016 and there was little difference between

1 annual average primary forest loss in 2000-2013 and in 2000-2014. Meanwhile, we
2 overlaid the map of wildfire-driven forest cover loss³³ with the maps of IFLs in 2000,
3 2013 and 2016, respectively, to estimate the annual average anthropogenic IFL loss in
4 2000-2013 and in 2000-2016 (we applied the same method for 2000-2013 instead of
5 using existing data for consistency). The difference between the two values is less than
6 0.05 Mha for most countries, which also supports our assumption. Detailed analyses of
7 the influence of amortization periods can be found in ref.⁴⁸.

8 Given these, an uncertainty analysis of the overall results has been performed to
9 reveal the uncertainties by adopting a stochastic modelling^{60,61}. The basic items *L*, *Z*
10 and *F* were perturbed 10000 times by introducing the standard deviation using Monte
11 Carlo simulation, based on which the perturbed demand-driven IFL loss can be obtained.
12 More technical details and source codes can be found in our previous work^{62,63}. The
13 relative standard deviation of the IFL inventory and GTAP MRIO used in this work
14 could be derived from Potapov et al.¹ and Hertwich et al.⁶⁴, respectively. The
15 uncertainties of the IFL loss embodied in final consumption varied from region to
16 region, from [-5.6%, +6.6%] for China to [-10.1%, +10.5%] for Brazil at the 95% level
17 of confidence (detailed results are presented in Table S4).

18
19

References

1. Potapov, P., Hansen, M.C., Laestadius, L., Turubanova, S., Yaroshenko, A., Thies, C., Smith, W., Zhuravleva, I., Komarova, A., Minnemeyer, S., and Esipova, E. (2017). The last frontiers of wilderness: Tracking loss of intact forest landscapes from 2000 to 2013. *Science Advances* 3, e1600821. 10.1126/sciadv.1600821.
2. Canadell, J.G., and Raupach, M.R. (2008). Managing Forests for Climate Change Mitigation. *Science* 320, 1456-1457. 10.1126/science.1155458.
3. Soares-Filho, B., Moutinho, P., Nepstad, D., Anderson, A., Rodrigues, H., Garcia, R., Dietzsch, L., Merry, F., Bowman, M., Hissa, L., et al. (2010). Role of Brazilian Amazon protected areas in climate change mitigation. *Proceedings of the National Academy of Sciences* 107, 10821. 10.1073/pnas.0913048107.
4. Barlow, J., Gardner, T.A., Araujo, I.S., Ávila-Pires, T.C., Bonaldo, A.B., Costa, J.E., Esposito, M.C., Ferreira, L.V., Hawes, J., Hernandez, M.I.M., et al. (2007). Quantifying the biodiversity value of tropical primary, secondary, and plantation forests. *Proceedings of the National Academy of Sciences* 104, 18555. 10.1073/pnas.0703333104.
5. Gibson, L., Lee, T.M., Koh, L.P., Brook, B.W., Gardner, T.A., Barlow, J., Peres, C.A., Bradshaw, C.J.A., Laurance, W.F., Lovejoy, T.E., and Sodhi, N.S. (2011). Primary forests are irreplaceable for sustaining tropical biodiversity. *Nature* 478, 378-381. 10.1038/nature10425.
6. Brown, J., Ferrians Jr, O.J., Heginbottom, J.A., and Melnikov, E.S. (1997). Circum-Arctic map of permafrost and ground-ice conditions. Report 45. <http://pubs.er.usgs.gov/publication/cp45>.
7. IFL Mapping Team. World's Intact Forest Landscapes 2000-2020. <https://intactforests.org/world.map.html>.
8. Pendrill, F., Gardner, T.A., Meyfroidt, P., Persson, U.M., Adams, J., Azevedo, T., Bastos Lima, M.G., Baumann, M., Curtis, P.G., De Sy, V., et al. (2022). Disentangling the numbers behind agriculture-driven tropical deforestation. *Science* 377, eabm9267. 10.1126/science.abm9267.
9. Meyfroidt, P., Börner, J., Garrett, R., Gardner, T., Godar, J., Kis-Katos, K., Soares-Filho, B.S., and Wunder, S. (2020). Focus on leakage and spillovers: informing land-use governance in a tele-coupled world. *Environmental Research Letters* 15, 090202. 10.1088/1748-9326/ab7397.
10. Liang, S., Guo, S., Newell, J.P., Qu, S., Feng, Y., Chiu, A.S.F., and Xu, M. (2016). Global Drivers of Russian Timber Harvest. *Journal of Industrial Ecology* 20, 515-525. 10.1111/jiec.12417.
11. Kan, S., Chen, B., Han, M., Hayat, T., Alsulami, H., and Chen, G. (2021). China's forest land use change in the globalized world economy: Foreign trade and unequal household consumption. *Land Use Policy* 103. 10.1016/j.landusepol.2021.105324.
12. Weinzettel, J., Hertwich, E.G., Peters, G.P., Steen-Olsen, K., and Galli, A.

- (2013). Affluence drives the global displacement of land use. *Global Environmental Change* 23, 433-438. 10.1016/j.gloenvcha.2012.12.010.
13. Hoang, N.T., and Kanemoto, K. (2021). Mapping the deforestation footprint of nations reveals growing threat to tropical forests. *Nature Ecology & Evolution*. 10.1038/s41559-021-01417-z.
 14. Pendrill, F., Persson, U.M., Godar, J., and Kastner, T. (2019). Deforestation displaced: trade in forest-risk commodities and the prospects for a global forest transition. *Environmental Research Letters* 14, 055003. 10.1088/1748-9326/ab0d41.
 15. Henders, S., Persson, U.M., and Kastner, T. (2015). Trading forests: land-use change and carbon emissions embodied in production and exports of forest-risk commodities. *Environmental Research Letters* 10, 125012. 10.1088/1748-9326/10/12/125012.
 16. Pendrill, F., Persson, U.M., Godar, J., Kastner, T., Moran, D., Schmidt, S., and Wood, R. (2019). Agricultural and forestry trade drives large share of tropical deforestation emissions. *Global Environmental Change* 56, 1-10. 10.1016/j.gloenvcha.2019.03.002.
 17. Fahrig, L. (2003). Effects of Habitat Fragmentation on Biodiversity. *Annual Review of Ecology, Evolution, and Systematics* 34, 487-515. 10.1146/annurev.ecolsys.34.011802.132419.
 18. Broadbent, E.N., Asner, G.P., Keller, M., Knapp, D.E., Oliveira, P.J.C., and Silva, J.N. (2008). Forest fragmentation and edge effects from deforestation and selective logging in the Brazilian Amazon. *Biological Conservation* 141, 1745-1757. 10.1016/j.biocon.2008.04.024.
 19. Veríssimo, A., Barreto, P., Tarifa, R., and Uhl, C. (1995). Extraction of a high-value natural resource in Amazonia: the case of mahogany. *Forest Ecology and Management* 72, 39-60. 10.1016/0378-1127(94)03432-V.
 20. Fearnside, P.M. (1997). Protection of mahogany: a catalytic species in the destruction of rain forests in the American tropics. *Environmental Conservation* 24, 303-306. 10.1017/S0376892997000404.
 21. Matricardi, E.A.T., Skole, D.L., Costa, O.B., Pedlowski, M.A., Samek, J.H., and Miguel, E.P. (2020). Long-term forest degradation surpasses deforestation in the Brazilian Amazon. *Science* 369, 1378. 10.1126/science.abb3021.
 22. Colin, A.P., Michael, S.Q., Pat, D., and Carla, P.C. (2016). The Landscape Impact of Linear Seismic Clearings for Oil and Gas Development in Boreal Forest. *Northwest Science* 90, 340-354. 10.3955/046.090.0312.
 23. Qin, Y., Xiao, X., Wigner, J.-P., Ciais, P., Brandt, M., Fan, L., Li, X., Crowell, S., Wu, X., Doughty, R., et al. (2021). Carbon loss from forest degradation exceeds that from deforestation in the Brazilian Amazon. *Nature Climate Change* 11, 442-448. 10.1038/s41558-021-01026-5.
 24. Kastner, T., Erb, K.-H., and Nonhebel, S. (2011). International wood trade and forest change: A global analysis. *Global Environmental Change* 21, 947-956.

- 10.1016/j.gloenvcha.2011.05.003.
25. Kan, S.Y., Chen, B., Wu, X.F., Chen, Z.M., and Chen, G.Q. (2019). Natural gas overview for world economy: From primary supply to final demand via global supply chains. *Energy Policy* 124, 215-225. 10.1016/j.enpol.2018.10.002.
 26. Wiedmann, T.O., Schandl, H., Lenzen, M., Moran, D., Suh, S., West, J., and Kanemoto, K. (2015). The material footprint of nations. *Proceedings of the National Academy of Sciences* 112, 6271-6276. doi:10.1073/pnas.1220362110.
 27. Aguiar, A., Chepeliev, M., Corong, E.L., McDougall, R., and van der Mensbrugge, D. (2019). The GTAP Data Base: Version 10. *Journal of Global Economic Analysis* 4, 1-27. 10.21642/JGEA.040101AF.
 28. Food and Agriculture Organization of the United Nations. FAOSTAT. Production and Trade. <https://www.fao.org/faostat/en/#data>.
 29. Olson, D.M., Dinerstein, E., Wikramanayake, E.D., Burgess, N.D., Powell, G.V.N., Underwood, E.C., D'Amico, J.A., Itoua, I., Strand, H.E., Morrison, J.C., et al. (2001). Terrestrial Ecoregions of the World: A New Map of Life on Earth: A new global map of terrestrial ecoregions provides an innovative tool for conserving biodiversity. *BioScience* 51, 933-938. 10.1641/0006-3568(2001)051[0933:TEOTWA]2.0.CO;2.
 30. Timber Trade Portal. Overview of Timber Sector of Democratic Republic of the Congo. <https://www.timbertradeportal.com/en/democratic-republic-of-the-congo/37/timber-sector>.
 31. Food and Agriculture Organization of the United Nations. FAO Yearbook of Forest Products 2014. <https://www.fao.org/3/i5542m/i5542m.pdf>.
 32. Cazzolla Gatti, R., Castaldi, S., Lindsell, J.A., Coomes, D.A., Marchetti, M., Maesano, M., Di Paola, A., Paparella, F., and Valentini, R. (2015). The impact of selective logging and clearcutting on forest structure, tree diversity and above-ground biomass of African tropical forests. *Ecological Research* 30, 119-132. 10.1007/s11284-014-1217-3.
 33. Curtis, P.G., Slay, C.M., Harris, N.L., Tyukavina, A., and Hansen, M.C. (2018). Classifying drivers of global forest loss. *Science* 361, 1108. 10.1126/science.aau3445.
 34. Garrett, R.D., Levy, S., Carlson, K.M., Gardner, T.A., Godar, J., Clapp, J., Dauvergne, P., Heilmayr, R., le Polain de Waroux, Y., Ayre, B., et al. (2019). Criteria for effective zero-deforestation commitments. *Global Environmental Change* 54, 135-147. 10.1016/j.gloenvcha.2018.11.003.
 35. World Bank Group. Minerals for Climate Action: The Mineral Intensity of the Clean Energy Transition. <http://pubdocs.worldbank.org/en/961711588875536384/Minerals-for-Climate-Action-The-Mineral-Intensity-of-the-Clean-Energy-Transition.pdf>.
 36. Forest Stewardship Council. Deforestation, high conservation value forests and intact forest landscapes. <https://fsc.org/en/deforestation-hcv-ifl>.
 37. Bager, S.L., Persson, U.M., and dos Reis, T.N.P. (2021). Eighty-six EU policy

- options for reducing imported deforestation. *One Earth* 4, 289-306. 10.1016/j.oneear.2021.01.011.
38. Hoare, A. (2015). Tackling Illegal Logging and the Related Trade - What Progress and Where Next. Chatham House, the Royal Institute of International Affairs.
<https://www.chathamhouse.org/sites/default/files/publications/research/20150715IllegalLoggingHoareFinal.pdf>.
 39. FAO and UNEP. The State of the World's Forests 2020. Forests, biodiversity and people. <http://www.fao.org/state-of-forests/en>.
 40. Gollin, D. (2014). Smallholder agriculture in Africa: An overview and implications for policy. IIED Working Paper. <https://www.iied.org/14640iied>.
 41. Fa, J.E., Watson, J.E.M., Leiper, I., Potapov, P., Evans, T.D., Burgess, N.D., Molnár, Z., Fernández-Llamazares, Á., Duncan, T., Wang, S., et al. (2020). Importance of Indigenous Peoples' lands for the conservation of Intact Forest Landscapes. *Frontiers in Ecology and the Environment* 18, 135-140. 10.1002/fee.2148.
 42. Maus, V., Giljum, S., Gutschlhofer, J., da Silva, D.M., Probst, M., Gass, S.L.B., Luckeneder, S., Lieber, M., and McCallum, I. (2020). A global-scale data set of mining areas. *Scientific Data* 7, 289. 10.1038/s41597-020-00624-w.
 43. Food and Agriculture Organization of the United Nations. FAOSTAT. Land Use. <https://www.fao.org/faostat/en/#data/RL>.
 44. Natural Resources Canada. Principal Mineral Areas, Producing Mines, and Oil and Gas Fields (900A). <https://open.canada.ca/data/en/dataset/000183ed-8864-42f0-ae43-c4313a860720>.
 45. Global Energy Observatory, Google, KTH Royal Institute of Technology in Stockholm, Enipedia, World Resources Institute. Global Power Plant Database. <https://datasets.wri.org/dataset/globalpowerplantdatabase>.
 46. ESA. Land Cover CCI Product User Guide Version 2. Tech. Rep. maps.elie.ucl.ac.be/CCI/viewer/download/ESACCI-LC-Ph2-PUGv2_2.0.pdf.
 47. Sonter, L.J., Herrera, D., Barrett, D.J., Galford, G.L., Moran, C.J., and Soares-Filho, B.S. (2017). Mining drives extensive deforestation in the Brazilian Amazon. *Nature Communications* 8, 1013. 10.1038/s41467-017-00557-w.
 48. Persson, U.M., Henders, S., and Cederberg, C. (2014). A method for calculating a land-use change carbon footprint (LUC-CFP) for agricultural commodities – applications to Brazilian beef and soy, Indonesian palm oil. *Global Change Biology* 20, 3482-3491. 10.1111/gcb.12635.
 49. Bhan, M., Gingrich, S., Roux, N., Le Noë, J., Kastner, T., Matej, S., Schwarzmüller, F., and Erb, K.-H. (2021). Quantifying and attributing land use-induced carbon emissions to biomass consumption: A critical assessment of existing approaches. *Journal of Environmental Management* 286, 112228. 10.1016/j.jenvman.2021.112228.
 50. British Standards Institute. PAS 2050:2011 Specification for the assessment of

- the life cycle greenhouse gas emissions of goods and services.
51. Zalles, V., Hansen Matthew, C., Potapov Peter, V., Parker, D., Stehman Stephen, V., Pickens Amy, H., Parente Leandro, L., Ferreira Laerte, G., Song, X.-P., Hernandez-Serna, A., and Kommareddy, I. Rapid expansion of human impact on natural land in South America since 1985. *Science Advances* 7, eabg1620. 10.1126/sciadv.abg1620.
 52. Kan, S., Chen, B., and Chen, G. (2019). Worldwide energy use across global supply chains: Decoupled from economic growth? *Appl. Energ.* 250, 1235-1245. 10.1016/j.apenergy.2019.05.104.
 53. Kan, S., Chen, B., Meng, J., and Chen, G. (2020). An extended overview of natural gas use embodied in world economy and supply chains: Policy implications from a time series analysis. *Energy Policy* 137, 111068. 10.1016/j.enpol.2019.111068.
 54. Davis, S.J., Peters, G.P., and Caldeira, K. (2011). The supply chain of CO₂ emissions. *Proc Natl Acad Sci USA* 108, 18554-18559. 10.1073/pnas.1107409108.
 55. Guan, D., Meng, J., Reiner, D.M., Zhang, N., Shan, Y., Mi, Z., Shao, S., Liu, Z., Zhang, Q., and Davis, S.J. (2018). Structural decline in China's CO₂ emissions through transitions in industry and energy systems. *Nature Geoscience* 11, 551-555. 10.1038/s41561-018-0161-1.
 56. Bullard, C.W., and Herendeen, R.A. (1975). The energy cost of goods and services. *Energy Policy* 3, 268-278. 10.1016/0301-4215(75)90035-X.
 57. Zhang, D., Caron, J., and Winchester, N. (2019). Sectoral Aggregation Error in the Accounting of Energy and Emissions Embodied in Trade and Consumption. *Journal of Industrial Ecology* 23, 402-411. 10.1111/jiec.12734.
 58. Bruckner, M., Fischer, G., Tramberend, S., and Giljum, S. (2015). Measuring telecouplings in the global land system: A review and comparative evaluation of land footprint accounting methods. *Ecological Economics* 114, 11-21. 10.1016/j.ecolecon.2015.03.008.
 59. University of Maryland and World Resources Institute. Global Primary Forest Loss. www.globalforestwatch.org.
 60. Lenzen, M., Sun, Y.-Y., Faturay, F., Ting, Y.-P., Geschke, A., and Malik, A. (2018). The carbon footprint of global tourism. *Nature Climate Change* 8, 522-528. 10.1038/s41558-018-0141-x.
 61. Zhang, H., He, K., Wang, X., and Hertwich, E.G. (2019). Tracing the Uncertain Chinese Mercury Footprint within the Global Supply Chain Using a Stochastic, Nested Input–Output Model. *Environmental Science & Technology* 53, 6814-6823. 10.1021/acs.est.8b06373.
 62. Wei, W., Li, J., Chen, B., Wang, M., Zhang, P., Guan, D., Meng, J., Qian, H., Cheng, Y., Kang, C., et al. (2021). Embodied greenhouse gas emissions from building China's large-scale power transmission infrastructure. *Nature Sustainability* 4, 739-747. 10.1038/s41893-021-00704-8.

63. Wei, W., Wang, M., Zhang, P., Chen, B., Guan, D., Shao, S., and Li, J. (2020). A 2015 inventory of embodied carbon emissions for Chinese power transmission infrastructure projects. *Scientific Data* 7, 318. 10.1038/s41597-020-00662-4.
64. Hertwich, E.G., and Peters, G.P. (2009). Carbon Footprint of Nations: A Global, Trade-Linked Analysis. *Environmental Science & Technology* 43, 6414-6420. 10.1021/es803496a.

Acknowledgements

The research is supported by the National Natural Science Foundation of China (Grant nos. 52100210 and 72061147003), the Swedish Research Council FORMAS (Grant nos. 213:2014-1181 and 2016-00351 (LEAKAGE)), the Deutsche Forschungsgemeinschaft (DFG, German Research Foundation) (Project No. KA 4815/1-1) and the German Federal Ministry for Economic Cooperation and Development (grant no. GS22 E1070-0060/029).

Author contributions

Siyi Kan: Conceptualization, Methodology, Software, Formal analysis, Data Curation, Visualization, Writing, Validation.

Bin Chen, U. Martin Persson and Thomas Kastner: Conceptualization, Methodology, Formal analysis, Writing, Validation, Funding acquisition.

Guoqian Chen: Conceptualization, Methodology, Supervision, Writing, Validation, Funding acquisition.

Yutao Wang, Jiashuo Li, Jing Meng, Heran Zheng, Lan Yang, Rui Li and Mingxi Du: Formal analysis, Writing, Validation.

Declaration of Competing interests

The authors declare no competing interests.

Figure captions

Figure 1 Loss of intact forest landscapes (IFLs) associated with final consumption.

Panels (A-C) break down IFL loss (A) by sector for major final consumers, (B) by sector for 6 region groups and (C) by percentage attribution to countries and regions. Panel (D) shows the links between proximate causes of IFL loss (left) and final consumption sectors (right). A full list of countries and regions is provided in Supplemental Information.

Figure 2 Loss of intact forest landscapes (IFLs) associated with international trade.

Panel (A) shows the displacement of IFL loss between the primary producers and final consumers. Only flows greater than 30 kha are marked. Panel (B) depicts source-to-sink budget of IFL loss. The left and middle columns represent IFL loss by regions and climate zones where it took place and the right column represents IFL loss embodied in final consumption by final consumers. “Rest of” regions on the left and right sides respectively refer to aggregations of all the countries not listed in corresponding columns.

Figure 3 Displaced loss of intact forest landscapes (IFLs) traced back to proximate causes.

Panels (A-B) further attribute displaced IFL loss to direct exploiters and final consumers, respectively. Panel (C) shows the share of displaced IFL loss in total loss caused by a specific economic activity. Panel (D) disaggregates IFL loss associated with final consumption by proximate causes and final consumers.

Figure 4 Regional contribution to production- and consumption-related intact forest landscape (IFL) loss.

Panel (A) reveals the influence of local consumption and export production on local IFL loss for dominant producer regions and Panel (B) compares the share of local consumption and import in total consumption-related IFL loss for dominant consumer regions.